Using Satellite Data to Fill the Gaps in the US Air Pollution Monitoring Network

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Abstract

Compliance with the Clean Air Act's National Ambient Air Quality Standards (NAAQS) is determined at the county level using local air pollution monitors. However, the vast majority of counties have zero or one monitor, and air pollution concentrations can vary dramatically over short distances. In addition, recent evidence suggests monitor placement may be skewed to make NAAQS compliance easier. As a result, the network of air pollution monitors may not detect all areas with pollution levels that exceed the NAAQS. This paper explores this possibility using satellite-derived data on fine particulate matter concentrations (PM_{2.5}) to assess NAAQS compliance for the continental United States at a 1 kilometer (km) resolution. We compare the satellite-based assessments with official attainment designations made by the terrestrial monitor network and calculate the number of people living in "misclassified" areas—areas with PM_{2.5} levels that exceed the NAAQS but are classified as being in attainment based on the ground monitors and the US Environmental Protection Agency's (EPA's) official determination. We estimate that about 24.4 million people are living in these misclassified areas about as many as are living in properly classified nonattainment areas—and that if such areas had sped up their PM_{2.5} reductions as much as nonattainment areas did, 5,452 premature deaths would have been avoided, a welfare gain to society of \$49 billion.

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1 Introduction

The Clean Air Act (CAA) is the foundation of air quality regulation in the United States. Under the CAA, air pollution monitors regularly measure concentrations of seven air pollutants ($PM_{2.5}$, PM_{10} , ozone, NO_x , SO_2 , CO, and lead) across the country. If a monitor reports a three-year average (a "design value") above the National Ambient Air Quality Standards (NAAQS) for a pollutant, the monitor's jurisdiction (usually the county) is classified as in "nonattainment" for that pollutant and is subject to a series of requirements to bring its design value down to the standard. Past research has shown that air quality improves significantly faster in nonattainment areas than in attainment areas, resulting in concomitant health and welfare benefits (see, e.g., Currie et al. 2014; Bishop, Ketcham, and Kuminoff 2018).

But air quality standards are only as good as the US Environmental Protection Agency's (EPA's) monitoring network, and this monitoring network is limited. The majority of US counties lack monitors altogether, and readings at an air pollution monitor do not necessarily represent concentrations across a wide area like a county. Of 3,100 counties in the United States, only 651 (21 percent) have any $PM_{2.5}$ monitors. Of these, 48 percent have a single monitor. In such cases, standard practice is to assume that the concentrations registered by that monitor are representative of concentrations throughout the county. Good monitor placement is obviously critical to this assumption, but recent research shows that some monitors appear to be placed in areas of low pollution relative to elsewhere in the county, such as upwind of major point sources (Grainger, Schreiber, and Chang 2018). 1

The purpose of this research is to measure how many people live in gaps in the monitoring network where pollution levels are high but undetected by monitors and therefore undetected by regulators. We use high-resolution satellite-derived data (\sim 1 km²) on ground-level PM_{2.5} concentrations to find counties that are designated as attainment but contain areas that the satellite data suggest violate the NAAQS.

We find that 54 counties in 11 states, home to 24.4 million people, are misclassified according to satellite data from the time period EPA used to originally assign attainment designations (2011–2013). Of these, 10.9 million live in counties that do not

^{1.} Auffhammer, Bento, and Lowe (2009) find that regulators strategically target certain monitors when implementing regulations, further highlighting the importance of monitor location.

contain any $PM_{2.5}$ monitors. Reclassifying all these misclassified individuals as nonattainment would more than double the total nonattainment population, currently 23.2 million.

One consequence of misclassification is that people living in misclassified counties did not enjoy the same accelerated improvements in air quality that properly classified nonattainment counties did after their nonattainment designation. Using a difference-in-differences approach, we estimate the decrease in $PM_{2.5}$ misclassified counties would have experienced had they been properly classified and use our estimates to calculate the number of premature deaths that would have been avoided. We find that improved air quality would have prevented 5,452 premature deaths. Using the value of a statistical life (VSL), this would result in \$49 billion in benefits to misclassified counties had states acted as quickly to reduce $PM_{2.5}$ levels in these areas as they have in nonattainment areas. This amount may be considered the value of remote sensing information in this air quality context.

2 Background on the Clean Air Act

The Clean Air Act (CAA) of 1970 and its subsequent amendments form the basis of current air quality regulation in the United States (Revesz 2015).²

The CAA directs the administrator of EPA to issue NAAQS for certain pollutants. Limits must be imposed on air pollutants that "may reasonably be anticipated to endanger public health or welfare" (42 USC § 7408(a)(1)(A)), and these limits should be set to "protect the public health" with "an adequate margin of safety" (§ 7409(b)(1)).³

The NAAQS are supposed to be reviewed and possibly revised no later than every five years, although this schedule is rarely met. These assessments are made in

^{2.} In general, see Revesz (2015), chapter 5, for detailed history and review of air quality regulation in the United States at the federal level.

^{3. § 7408(}a)(1)(B) also requires that the presence of the pollutant be due to "numerous or diverse mobile or stationary sources." The public health standard prescribed in § 7409(b)(1) is known as the "primary" standard. § 7409(b)(2) provides for "secondary" standards that "protect the public welfare." Welfare "includes, but is not limited to, effects on soils, water, crops, vegetation, manmade materials, animals, wildlife, weather, visibility, and climate, damage to and deterioration of property, and hazards to transportation, as well as effects on economic values and on personal comfort and well-being" (§ 7602(h)). Note, however, that the costs of meeting the standards may not be considered when setting standards.

consultation with an independent scientific review committee, the Clean Air Scientific Advisory Committee (CASAC), which provides scientific and technical advice.

The current primary standards for $PM_{2.5}$ were set in 2012 as (1) an annual average of no more than 12 micrograms per cubic meter ($\mu g/m^3$) (down from 15 $\mu g/m^3$ in 2005) and (2) a 98th percentile of daily readings no more than 35 $\mu g/m^3$ (unchanged from 2005). Both metrics are calculated using the three most recent years of monitor data. As we discuss further in Section 5, this study focuses on the annual standard of 12 $\mu g/m^3$.

Once a NAAQS is established for a pollutant, each state formally recommends to EPA which areas (generally counties) should be classified as in attainment with the NAAQS, which should be classified as nonattainment areas (§ 107(d)(1)), and which are nonclassifiable (effectively attainment) areas. The state is required to use the latest three years of monitoring data to do this, but it also may use atmospheric modeling, emissions inventories, and other tools.⁴ States also identify areas that contribute to downwind air quality violations and include them in their nonattainment recommendations. EPA examines the state submission and then makes its determination on nonattainment designations, which states can appeal.

The CAA also permits reclassifications of an area's status as conditions change. Most reclassifications are made after states or other groups petition to move from nonattainment status to attainment. EPA can also reclassify in the other direction on its own or if asked to by petitioners, but this rarely happens. EPA data show that only one area has ever been reclassified from attainment status to nonattainment status for any $PM_{2.5}$ standard.⁵

Once an area is officially designated nonattainment, the state or states in which the area is located must submit a state implementation plan (SIP) to EPA that outlines how the NAAQS will be met (e.g., what restrictions will be placed on which industries in which parts of the state). Polluters in nonattainment areas face more stringent regulations than those in attainment, such as a requirement to use the best available

^{4.} When the $PM_{2.5}$ standards were set in 2012, EPA directed states to use air quality monitoring data from 2010 through 2012 in their initial recommendations for nonattainment areas and said it would use data from 2011 through 2013 in its final determination if 2013 data were available in time.

^{5.} Pinal County, Arizona, was reclassified in 2011 as nonattainment with the $PM_{2.5}$ 2006 rule (24-hour standard) two years after initial classifications were made for that rule in 2009. Reclassification from attainment to nonattainment for any pollutant is rare. It has happened only 59 times across all 13 criteria pollutant standards. Of these, 34 percent were changes to counties' SO_2 status (2010 rule) and occurred in the last three years.

control technology. Areas with more severe nonattainment designations face tighter restrictions but have longer deadlines to reach attainment. States that continually fail to make "reasonable further progress" in reaching attainment may face federal funding sanctions or other consequences. Understanding the schedule and speed of nonattainment areas in reaching attainment is important when we estimate the health benefits of making proper designations.

3 Problems with a Limited Network of Air Pollution Monitors

The NAAQS and the attainment designations depend on EPA's network of pollution monitors to provide an accurate measure of how much pollution people are exposed to. However, there are several problems with using a limited network of air pollution monitors to measure exposure of a spatially dispersed population. The fundamental issues are that (1) air pollution varies significantly over short distances, with spikes around every factory, every road, and every refinery; 6 (2) pollution can travel long distances from its source; and (3) accurate monitors are expensive to buy and operate. In 2015, 79 percent of counties did not have a $PM_{2.5}$ monitor, 10 percent had one monitor, 5 percent had two monitors, and 6 percent had three or more. Even if a county has a monitor, it may not operate very often. For example, 56 percent of $PM_{2.5}$ monitors gathered data on fewer than 121 days in 2015, and 23 percent of those gathered data on fewer than 80 days.

In addition to physical problems of measuring pollution, there are two problems with regulatory air pollution monitors that arise because air pollution is generated by economic activity. The first problem is Goodhart's law: once a metric of economic activity becomes a regulatory target, it is no longer a good metric of the underlying activity. This is because economic actors may adapt their behavior to affect the metric, like when a teacher changes his curriculum to better fit standardized tests used to evaluate his performance. For air pollution, this "teaching to the test" could happen in a number of ways: strategic placement of monitors in less-polluted parts of the county, strategic timing of abatement by polluters when monitors are in operation, or the

^{6.} For ozone, there may be reductions in concentrations close to sources of the ozone precursors nitrogen oxides (NO_x) and volatile organic compounds (VOCs) (termed scavenging), such as when NO_x emissions are decreased where VOCs are limiting ozone formation.

gradual relocation of polluters over time from locations upwind of monitors to locations downwind. Grainger, Schreiber, and Chang (2018) present evidence of strategic monitor placement. Zou (2017) presents evidence that firms reduce their pollution on days that PM_{2.5} monitors are in operation, since these monitors operate only on select days and their schedule is published in advance. Our paper does not address these problems directly, but our results are likely driven in part by them. The second problem is that monitors are fixed in space and time, while the location and timing of pollution are constantly changing. A pollution monitor provides a sample concentration from a single point in what could be a large area with varying topography, wind conditions, traffic patterns, and density of industry. Since air quality regulation is ultimately aimed at improving health, how people, pollution, and polluters are differentially distributed across space cannot be ignored. Furthermore, none of these distributions are static. As polluters in different locations change their polluting behavior, as establishments relocate or open for the first time, the spatial distribution of emissions changes, and in turn the overall exposure to nearby residents changes. As neighborhoods grow in some parts of a city and shrink in others, overall exposure changes. As the local climate becomes hotter, airborne pollutants react in different ways and total exposure changes. Meanwhile, the monitor observes the air at the same physical location and is blind to all these changes. If a monitor initially corresponds to the median concentration in a county, a year later it may correspond to the 40th percentile, the 60th percentile, or some other order statistic from the population exposure distribution.

These problems together motivate a closer look at how well the monitors in the United States measure resident exposure to air pollution and, in turn, at whether the process of designating areas can be improved through the use of remote sensing data.⁷

4 Research Design

Are there areas exceeding the NAAQS that EPA's monitor network has missed? What improvements in pollution-related mortality would occur if these areas engaged in the

^{7.} Our focus is on areas misclassified as being in attainment that actually have $PM_{2.5}$ concentrations exceeding the NAAQS. Readers may naturally ask, What about areas classified as nonattainment that actually have readings below the standard? At the county level, this is an empty set—no counties are wrongly classified in this way. However, within a given county, there of course are areas shown by the monitors and, with much greater resolution, by the satellite data to have concentrations below the standard. We plan to take up this issue in terms of the appropriate spatial definition of nonattainment areas (particularly in light of the availability of satellite data) in a future paper.

same PM_{2.5} mitigation efforts employed by actual nonattainment areas?

To answer these questions, we first compare the satellite-derived data on ground-level $PM_{2.5}$ with official nonattainment designations to flag census blocks which (1) have satellite $PM_{2.5}$ readings over the NAAQS and (2) are classified as attainment/unclassifiable. We refer to counties with any such areas as "misclassified," since they would have been classified as nonattainment had EPA's monitor network had the same spatial coverage as the satellite data.

We then estimate the excess mortality in misclassified counties. A nonattainment designation pushes local regulators to lower pollution in order to get their county to attainment status. Past research has found that regulators can be effective at targeting monitors that cause their county to be in nonattainment. As Auffhammer, Bento, and Lowe (2009) observe, regulators in nonattainment areas have less incentive to target monitors that are not over the NAAQS because the nonattainment designation depends on readings at the highest monitor. We take an empirical approach similar to that of Auffhammer, Bento, and Lowe (2009); Grainger (2012); and Bento, Freedman, and Lang (2015) to estimate the effect of $PM_{2.5}$ nonattainment status on violating monitors. In particular, we estimate a difference-in-differences regression to measure the effect on $PM_{2.5}$ concentrations over time for monitors in nonattainment areas that register readings over the NAAQS (termed Group I) versus monitors in nonattainment areas that register concentrations below the NAAQS (termed Group II) versus monitor readings in attainment areas (termed Group III):

$$P_{mt} = \beta_1 \left(\text{Nonattainment}_m \times \text{post}_t \times \text{OverNAAQS}_m \right) +$$

$$\beta_2 \left(\text{Nonattainment}_m \times \text{post}_t \right) +$$

$$\delta_t + \delta_m + \varepsilon_{mt}$$

$$(1)$$

where P_{mt} is the pollution reading for monitor m at time t. The indicator variables δ_t and δ_m control for year and monitor effects. Nonattainment and "over NAAQS" status are taken from the year 2015, the first year nonattainment determinations were made

^{8. &}quot;The federal regulation creates an incentive for the local regulator to closely track the monitors that put the county at 'risk' of becoming out of attainment. The regulator then allocates effort in terms of monitoring and enforcement activities to the different monitors by comparing the future costs of getting out of attainment to the present costs associated with the reduction in the emissions around 'risky' monitors. The resulting equilibrium is a schedule of heterogeneous monitoring and enforcement efforts such that more effort is allocated to dirtier monitors" (Auffhammer, Bento, and Lowe 2009, 17).

for the 2012 PM_{2.5} rule. Here β_2 is the base effect of being a nonattainment monitor (Group I and Group II) and β_1 is the additional effect of being a monitor targeted for reductions by local regulators (Group I). Thus, the impact of nonattainment status on non-targeted nonattainment monitors (Group II) is β_2 , and the impact on targeted nonattainment monitors (Group I) is $\beta_1 + \beta_2$.

After estimating $\hat{\beta}_1$ and $\hat{\beta}_2$, we consider a scenario with spatially dense monitoring (e.g., satellite-derived data at 1 km resolution) in which regulators target all areas that exceed the NAAQS as opposed to one or two particular locations, as is the case with sparse monitoring. We assume that if misclassified counties had been correctly classified, they would have experienced declines in $PM_{2.5}$ similar to those experienced in counties actually designated as nonattainment. We then use the concentration-response estimate from Lepeule et al. (2012) to calculate excess mortality—that is, how many deaths would have been avoided if misclassified counties had been classified as nonattainment.

We focus on the most recent revision to the $PM_{2.5}$ NAAQS, which was made in 2012, lowering the limit for annual average $PM_{2.5}$ to $12\,\mu g/m^3$. States submitted recommendations for their nonattainment designations in 2014 using monitor data from 2011 to 2013, and official designations were announced in 2015.

5 Data

5.1 Air pollution monitors and attainment designations

Data on EPA's air pollution monitoring system come from EPA and cover every air pollution monitor from 1999 through 2017. The data include latitude and longitude, days of operation, pollution readings, and whether the monitor can be used to determine NAAQS compliance. When used, average annual readings for each monitor exclude concurred exceptional events.⁹

Table 1 reports the number of monitors available for NAAQS compliance during our period of study. Panel A reports the number of monitors designated as NAAQS primary

^{9.} In exceptional events that are outside state regulators' control, such as wildfires, state regulators may petition to have monitor readings from those events excluded from design value averages. EPA then chooses whether to concur that the event was exceptional and allow it to be excluded from the design value.

compliance monitors that operated in the given year. Column 1 reports how many monitors operated no more than 80 days during that year, column 2 reports how many operated 81–120 days, and so on. The strongest time trend is the addition of monitors operating more than 300 days a year. Panel B reports how many monitors had sufficient data over the prior three years to calculate a design value. For any given year and frequency, the number of monitors with three years of data is generally less than those with one year of data, though small anomalies can occur when monitors move across the frequency categories year to year. Both panels show that a significant proportion of monitors operate less than once every three days (columns 1 and 2). Even as late as 2017, 10 percent of monitors available to calculate design values to compare with the NAAQS operated no more than 80 days per year.

Data on attainment status are taken from EPA's Green Book.¹⁰

5.2 Satellite-derived concentration data

The satellite-derived PM_{2.5} concentration data come from a variety of sources (van Donkelaar et al. 2015; van Donkelaar et al. 2016). The data are primarily gathered by satellite-based instruments that measure aerosol optical depth (AOD). The best known of these instruments among economists are the MODIS instruments aboard the Terra and Aqua satellites (see, e.g., Zou 2017; Grainger, Schreiber, and Chang 2018; Gendron-Carrier et al. 2018). As these satellites orbit Earth, the MODIS instruments on board capture data on the density of airborne particles. It does this by comparing the intensity of solar radiation at the top of the atmosphere with how much radiation is reflected by Earth's surface. The more airborne particles there are to scatter and absorb this radiation, the less radiation is reflected to the satellite.

Both satellites follow a polar orbit, going from the North Pole to the South Pole and back to the North Pole every 100 minutes or so. As the satellites orbit pole to pole, Earth continues to rotate, giving the satellites a new swath of ground to scan. The satellites' orbits are calculated so that they pass over and scan any given point on Earth at approximately the same time every day. On the sun-facing side of Earth, Terra crosses the equator at approximately 10:30 a.m. local time with each orbital pass, while Aqua

^{10.} See https://www.epa.gov/green-book/green-book-data-download.

^{11.} The data we use here are an updated version of the North America data developed in van Donkelaar et al. (2015), which uses advances presented in van Donkelaar et al. (2016). We are very grateful to Aaron van Donkelaar for giving us access to these data.

crosses the equator at approximately 1:30 p.m. Thus, every location is scanned by each satellite approximately once per day at roughly the same time every day. These once-a-day readings are temporally sparser than hourly readings available from ground monitors. However, as discussed in the previous section, few monitors report hourly data, and most do not collect data every day.

Van Donkelaar et al. (2015) and van Donkelaar et al. (2016) combine the AOD data from the MODIS MISR (also aboard Terra) and SeaWIFS (aboard OrbView-2 satellite) instruments with results from the chemical transport model GEOS-Chem. GEOS-Chem provides information about how pollutants are transported from one area to another by the wind and how chemical compounds change as they travel. This combination of measurements and simulation is calibrated using ground-based monitored observations of $PM_{2.5}$ at a monthly timescale. The data are then averaged by year for every 0.01-by-0.01-degree grid cell, which is approximately 1 km² in area.

While satellite-derived data on air pollution provide unique opportunities for researchers and policy makers, they also come with a few caveats. First, the satellites do not measure $PM_{2.5}$ directly. They measure AOD, which must be scaled to $PM_{2.5}$ based on local conditions. This is not altogether straightforward even for researchers with atmospheric sciences training, but AOD itself can sometimes be used to estimate comparative $PM_{2.5}$ levels in localized areas. However, when comparing data with general policy thresholds such as the NAAQS, they must be accurately scaled. Second, satellites cannot measure ground-level conditions on cloudy days. This is one reason the satellite-derived data are more reliable at large timescales (months or years) than small ones (hours or days). Third, the accuracy of the data depends on the sample of ground-based monitors used for calibration. For example, data calibrated globally could have mean-zero error globally, but sub-samples of the data (e.g., the data for North America) may not be mean zero. To avoid this problem, we use data specifically calibrated for North America, which are quite accurate. 12

Figure 1 plots the correlation between the van Donkelaar et al. satellite-derived data (vertical axis) and annual average readings from ground-based monitors (horizontal axis). The satellite data for each monitor is taken from the 0.01-by-0.01-degree cell in

^{12.} Compare correlation between monitors and satellite data calibrated for North America shown in Figure 1, discussed below, and the equivalent figure for globally calibrated data restricted to North America in Figure A1, which shows a systematic upward bias relative to the monitors.

which the monitor is located. Faint markers indicate individual monitor–grid cell pairs; bold markers indicate the average for every bin centered at integers on the horizontal axis (i.e., satellite average for monitor readings of $1\pm0.5~\mu g/m^3$). The shape and color of each marker indicates how frequently the monitor operates: red circles for monitors that operate no more than 80 days per year; yellow triangles for 81–120 days; blue squares for 121–300 days; and green pentagons for those operating more than 300 days per year. Dashed gray lines show the 12 $\mu g/m^3$ NAAQS threshold for nonattainment classification. In general, the satellites show strong agreement with the monitors, especially at lower monitor readings. At higher monitor readings, the satellites tend to underestimate pollution concentrations relative to the monitors. This would imply that our methodology may be somewhat conservative in determining areas that are misclassified as attainment.

5.3 Population and demographic data

Block-level data on population counts and race/ethnicity come from the 2010 census. Block group–level data on educational attainment and household income come from the 2005–2010 American Community Survey (ACS). Data on county-level all-cause mortality come from the Centers for Disease Control and Prevention's (CDC's) Compressed Mortality File. ¹³

6 Results

6.1 Monitor coverage and nonattainment status

We begin by looking at the locations of $PM_{2.5}$ monitors in the continental United States. Figure 2 shows the location of each of the monitors that were used to make the 2015 attainment determinations. It also labels monitors based on how many days per year the monitor is required to operate following the same scheme as in Figure 1. We see significant heterogeneity across states in both the density of monitors and the frequency of their use. Some states have dense monitor networks that operate daily or near daily (e.g., California, Pennsylvania). Others are hardly monitored at all (e.g.,

^{13.} See https://wonder.cdc.gov/cmf-icd10.html.

Montana, Maine, Mississippi, Nebraska, Nevada, Idaho).¹⁴ Still others have many monitors, but each of those monitors does not operate more than 80 days per year (e.g., Wisconsin, Wyoming). Even within states, coverage can vary. Most of California is densely monitored, but in central California the monitors operate nearly every day, while in the Los Angeles Basin they operate no more than once every six days.

Figure 3 shows the designated nonattainment areas established in 2015 for the annual $PM_{2.5}$ standard promulgated in 2012. These nonattainment areas cover central California; the Los Angeles Basin; West Silver Valley, Idaho; Cleveland; Pittsburgh; and Philadelphia. All these areas were designated as nonattainment because of high monitor readings and not because they contributed to a downwind nonattainment area.

The key question the satellite data can answer is whether other counties also exceed the NAAQS.

6.2 Satellite-measured concentrations and misclassified areas

Figure 4 shows the difference between the satellite-measured three-year average $PM_{2.5}$ concentrations and the annual NAAQS for the continental United States in 2015. Blue areas correspond to those for which the satellite predicts a design value below the NAAQS ($12\,\mu g/m^3$), red areas are those predicted above the NAAQS, and white areas are those right at the NAAQS. This map shows many areas above the NAAQS, particularly in California, where Figure 3 also showed large nonattainment areas. However, Figure 4 also suggests that a large share of the Midwest is close to the NAAQS, and several areas have concentrations well above the NAAQS despite being classified as attainment areas.

Figure 5 focuses on some of these hot spots, highlighting the area bounded by Chicago on the north and west, Louisville on the south, and Pittsburgh on the east. Again, red corresponds to concentrations over the NAAQS, white about equal to the NAAQS, and blue under the NAAQS. County boundaries are plotted in white, official nonattainment areas are bounded in orange, and monitor locations are represented by black dots.

This map gives several examples of misclassified areas. First are areas with concentrations exceeding the NAAQS but that have no monitors in their counties (e.g.,

^{14.} Illinois is a special case where the monitors present in the state were deemed by EPA to be of insufficient quality to be used for NAAQS assessment. Therefore, all the data were thrown out, and the entire state was classified as attainment/unclassifiable.

Logansport, Indiana, north of Indianapolis and southeast of Chicago). Second are areas exceeding the NAAQS that are not detected because the monitors are too far away from the hot spots (e.g., southeast of Logansport). Third are cities with multiple monitors, but those monitors are located on the edges of the hot spots and miss peak concentrations (e.g., Indianapolis, Louisville, and Cincinnati).

Figure 6 shows all the misclassified counties—those that were designated as attainment but contain areas that exceed the NAAQS—in the continental United States. There are 54 such counties across 11 states. ¹⁵ Table 2 lists the number of people living in misclassified counties in each state, with separate counts for counties with monitors and without. For counties that include areas that are both attainment and nonattainment, we treat the attainment part of the county as a distinct county. 16 All told, 24.4 million people live in misclassified areas. Of these, 10.9 million live in counties with no monitors. The states with the largest populations of unmonitored and misclassified people are Illinois (6.4 million misclassified, all unmonitored); California (4.9 million misclassified, of which 0.8 million are unmonitored); and Texas (4.5 million misclassified, with 0.4 million unmonitored). Two other states have sizable misclassified populations that are unmonitored: Kentucky (1.2 million misclassified, 1 million unmonitored) and Ohio (2.2 million misclassified, 0.9 million unmonitored). The total number of people living in misclassified counties is 24.4 million, slightly more than the number of people currently living in official nonattainment areas (23.3 million).17

6.3 Population characteristics by attainment status

Table 3 summarizes the demographics of correctly classified attainment areas, misclassified areas, and nonattainment areas using 2010 census block-level data on population counts, race/ethnicity, and share urban; and 2005–2010 American Community Survey (ACS) block group–level data on education and income. The 23.2

^{15.} Table A1 lists all misclassified counties and their core-based statistical areas.

^{16.} If the satellite data find that the attainment portion should also be nonattainment, we count only people living in the attainment portion as misclassified. Similarly, if the nonattainment portion has pollution monitors but the misclassified attainment portion does not, we describe the attainment "county" as having no monitors.

^{17.} We include in these counts Chicago (and the rest of Illinois) and Houston whose designation is attainment/unclassifiable because their monitoring was not deemed reliable enough to determine NAAQS compliance.

million people living in designated nonattainment areas are more likely to live in an urban environment than residents in properly classified attainment areas (97 percent vs. 78 percent). They are also less likely to be white (non-Hispanic) than attainment residents (39 percent vs. 67 percent), slightly less likely to be black (9 percent vs. 13 percent), and much more likely to be Hispanic (40 percent vs. 14 percent). Household incomes tend to be slightly higher in nonattainment areas versus attainment areas. However, nonattainment residents are less likely to have finished high school (21 percent without a diploma vs. 14 percent), while equal shares of residents in both areas attended and completed college.

The racial makeup of misclassified areas is slightly different than that of nonattainment areas. Residents in the former are more likely to be white and more likely to be black, but less likely to be Hispanic or Asian. Income and educational attainment are virtually indistinguishable between the two areas. Thus, there does not appear to be any distinct demographic pattern to misclassified areas, with any distinction likely being driven by differences in urban demographics in different states (e.g., Indianapolis and Louisville vs. Los Angeles).¹⁸

6.4 Excess deaths from being misclassified as attainment

Nonattainment classification improves air quality because regulators are required to develop and implement plans to progress toward attainment. Had the misclassified areas listed above instead been designated as nonattainment, we assume their regulators would have acted the same way and the 24.4 million residents of these areas would have enjoyed cleaner air and had lower risk for mortality and various morbidities.

Accordingly, we estimate the excess mortality that could have been avoided in misclassified areas by measuring the effect nonattainment designation had on the air quality in actual nonattainment areas versus attainment areas. We then suppose that misclassified areas would have seen similar improvements had they been properly classified and calculate how many deaths might have been avoided by using $PM_{2.5}$ concentration-mortality response estimates from Lepeule et al. (2012). Different

^{18.} Because of the large sample sizes used, all but two of the means in column 2 of Table 3 are statistically significantly different with p-values less than 0.001. Those two variables are "fraction of households with income over \$75,000" and "fraction with some college."

^{19.} See Chay and Greenstone (2003), Currie et al. (2014), and Bishop, Ketcham, and Kuminoff (2018).

estimates of mortality avoidance would be generated by using concentration-response functions from other studies. Thus, these estimates are meant to be illustrative. However, numerous EPA Regulatory Impact Analyses, including one for $PM_{2.5}$ standards (EPA 2012), feature estimates from Lepeule et al. (see also, e.g., EPA 2014; 2015).

Figure 7 plots the average readings of the three groups of monitors introduced above: Group I is those monitors in nonattainment areas with readings over the NAAQS; Group II is those in nonattainment areas with readings under the NAAQS; and Group III is those in attainment areas. Both Group II and III monitors show steadily decreasing pollution levels over time, with a slight downward break in trend after nonattainment designations are made in 2015. The Group I monitors show a marked drop in pollution following the 2015 designations, going from 15 μ g/m³ to just under 10 μ g/m³. This much larger drop and the roughly parallel trends in the other two groups of monitors are consistent with the finding that regulators target violating monitors in particular. The difference-in-differences regression described in Section 4 formally estimates the impacts of a nonattainment classification on Group I and Group II monitors relative to Group III.

Table 4 reports the regression results. The regression sample is every $PM_{2.5}$ monitor in operation continuously from 2013 to 2017. Column 1 reports the regression of monitor readings on year indicators and a nonattainment–post interaction. The "post" period is 2016 and 2017, following the nonattainment designations made in 2015. The preferred specification in column 2 adds a variable identifying monitors in nonattainment areas operating in 2016–2017 and showing violation of the NAAQS (nonattainment–post–over NAAQS), a triple interaction term. Both regressions also account for any idiosyncrasies attributable to specific monitors through monitor-level fixed effects. All standard errors are clustered by monitor.

Column 1 shows that the average monitor in nonattainment areas records $1.14\,\mu g/m^3$ less $PM_{2.5}$ after the 2015 nonattainment designations relative to monitors in attainment areas. Column 2 allows for a separate effect on monitors over the NAAQS and shows that the overall decrease in pollution is being driven by monitors over the NAAQS.

^{20.} Here we really mean "continuously according to schedule," so a monitor that was supposed to operate every sixth day is considered to have "continuously" operated if it did so for the entire period in question.

Under-NAAQS nonattainment monitors see pollution drop $0.47 \,\mu\text{g/m}^3$ relative to attainment monitors, while over-NAAQS monitors see an additional $2.30 \,\mu\text{g/m}^3$ decrease, for a total decrease relative to attainment monitors of $2.77 \,\mu\text{g/m}^3$.

To calculate excess mortality from misclassification, we suppose that if misclassified areas had been designated as nonattainment, they would have experienced an average decline in pollution levels similar to that in areas properly designated as nonattainment. Specifically, areas over the NAAQS would have their $PM_{2.5}$ decrease by an additional 2.77 $\mu g/m^3$, while $PM_{2.5}$ in non-exceeding areas in the same county would decrease by 0.47 $\mu g/m^3$. We translate these pollution decreases into decreased mortality risk by multiplying them by the concentration-response coefficient from Lepeule et al. (2012) of 14 percent increase in all-cause mortality per additional 10 $\mu g/m^3$ $PM_{2.5}$. This leads to a 3.9 percent increase in all-cause mortality in areas over the NAAQS and a 0.7 percent increase in mortality in the rest of the county. Finally, we multiply these figures by county-specific death rates from the CDC and block-level population from the census.

We find that misclassified counties would have avoided 2,726 premature deaths per year had they been correctly classified. Using EPA's standard VSL of \$9 million, the social cost of this excess mortality is approximately \$24.5 billion per year (EPA 2016, 4-16). While the excess mortality effect is measured annually, it eventually gets eliminated as pollution trends in nonattainment areas equalize with those in attainment areas. A conservative assumption would be that the trends shown in Figure 7 will equalize after 2017 (after our regression sample) so that all benefits are realized in 2016 and 2017. This would imply that total excess mortality of misclassification was 5,452, with a social cost of \$49 billion.

7 Conclusion

The Clean Air Act is the primary air quality regulation in the United States. However, its success in improving health and environmental quality depends on a limited network of stationary pollution monitors to provide regulators with information about local pollution levels. If pollution levels in an area exceed the NAAQS but there is no monitor nearby, that area is unlikely to exercise mitigation actions to reduce its pollution. In this paper, we have used satellite data to provide evidence that significant portions of the country are indeed exceeding the annual $PM_{2.5}$ NAAQS standard but nevertheless

are designated as being in attainment. Then we have shown that if regulators in these misclassified areas acted to reduce pollution in the same way as regulators in properly designated nonattainment areas, thousands of lives could have been saved by their reclassification, a potential welfare gain to society of almost \$50 billion.

While the value to health and social welfare provided by satellite information on air quality appears to be very large, a few caveats are in order. The main caveat is that while satellite data are far more spatially dense than ground-based monitoring data, the conversion of what is actually measured by the satellites (aerosol optical depth) to $PM_{2.5}$ is not without error or bias when compared with monitor readings at the same place. In our case, the bias works to make our surprisingly large estimates of misclassifications conservative. The other caveat is the relative temporal sparseness of reliable satellite data. To achieve high spatial resolution (while maintaining accuracy) requires aggregating to larger time scales. Yet, on a more macro scale, the satellites provide data for every day, while in 2016 at most 37 percent of monitors were operating daily.

When the CAA first became law in 1970, legislators could not have envisioned the capability of measuring air quality on a spatially precise basis from satellites. Our results suggest that EPA should examine whether there is scope for reclassifying areas according to satellite information, or at least using satellite data as one of several factors that enter into the designation decision. Failing that, the Clean Air Act should be reopened to change the designation process to better protect the health of the US population.

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Figures and Tables

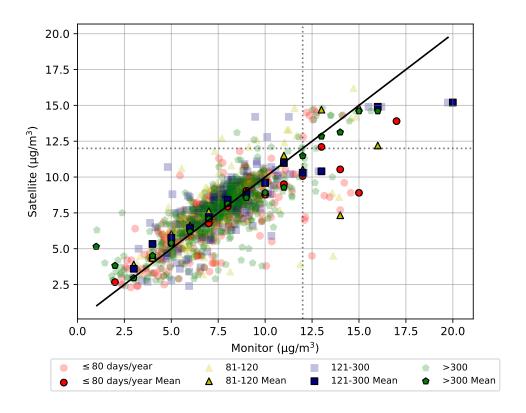


Figure 1: Satellite Readings versus Monitor Readings

Notes: Horizontal axis is annual average monitor reading. Vertical axis is the satellite-derived reading for the 0.01-by-0.01-degree cell where the monitor is located. Red circles indicate monitors that operate no more than 80 days per year; yellow indicate triangles 81–120 days; blue squares indicate 121–300 days; and green pentagons indicate more than 300 days per year. Faint markers indicate individual marker–grid cell pairs; bold markers indicate the average for every bin centered at integers on the horizontal axis, i.e., satellite average for monitor readings of $1 \pm 0.5 \,\mu\text{g/m}^3$. Dashed gray lines show the $12 \,\mu\text{g/m}^3$ NAAQS threshold for nonattainment classification.

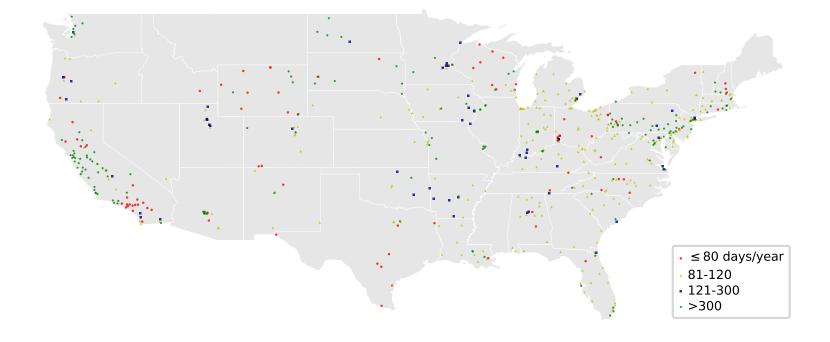


Figure 2: PM_{2.5} Monitors by Temporal Coverage, 2011–2013

Notes: Categories determined using median of valid observation days from 2011 to 2013. Monitors must cover entire time period to be included in sample. Red dots denote monitors that operate no more than 80 days per year; yellow triangles denote 81–120 days per year; blue squares denote 121–300 days; and green pentagons enote at least 300 days per year.



Figure 3: Clean Air Act Attainment Status, 2015 Notes: Darker areas are those classified as nonattainment with $PM_{2.5}$ 2012 primary standard of 12 $\mu g/m^3$.

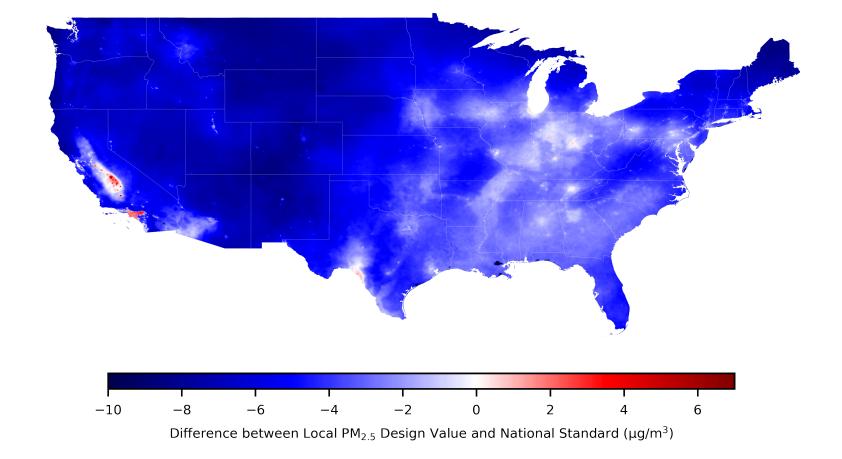


Figure 4: Satellite-measured $PM_{2.5}$ Design Values, 2015 Notes: $PM_{2.5}$ design values come from the satellite data described in Section 5. Plotted concentration is 3-year lagged average (2011–2013), which is the design value used to measure compliance with the NAAQS.

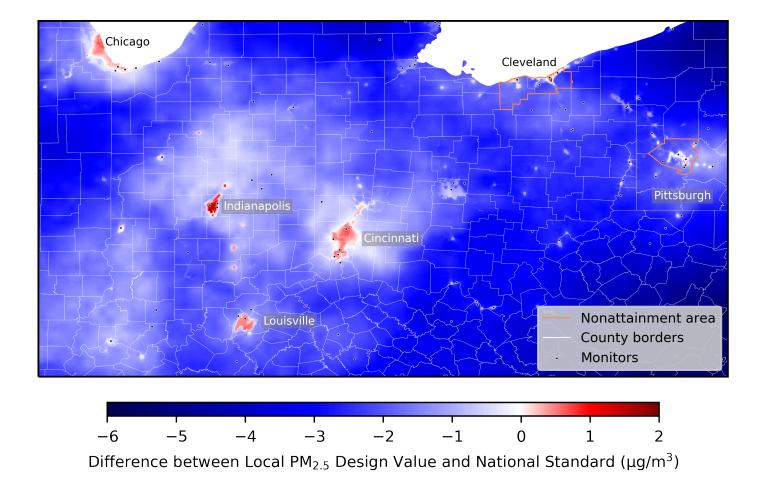


Figure 5: PM_{2.5} Design Values and Attainment Status

Notes: Orange boundaries indicate official nonattainment areas for the $PM_{2.5}$ 2012 primary standard of $12 \,\mu g/m^3$. Plotted $PM_{2.5}$ design values come from the satellite data described in Section 5 and are the average of years 2011–2013, the years of data that were used in making 2012 rule determinations. Monitor sample restricted to those used for NAAQS assessments.

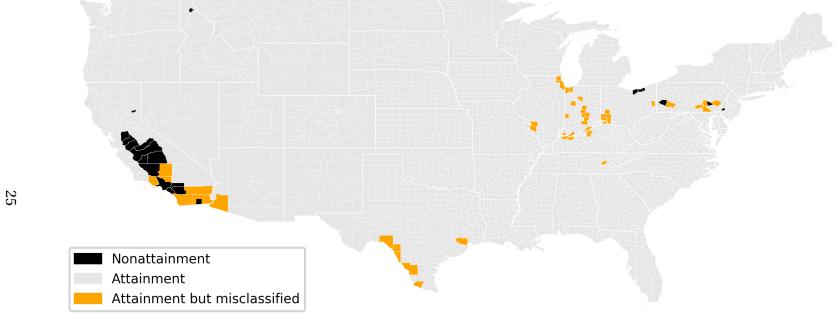


Figure 6: Areas Misclassified as Attainment for PM_{2.5} Annual Standard

Notes: Black areas denote official nonattainment counties and sub-counties. Yellow areas are counties that are misclassified, i.e., counties that are officially attainment where the satellite data show that some portion of the county exceeds the NAAQS. Gray areas are correctly classified attainment counties.

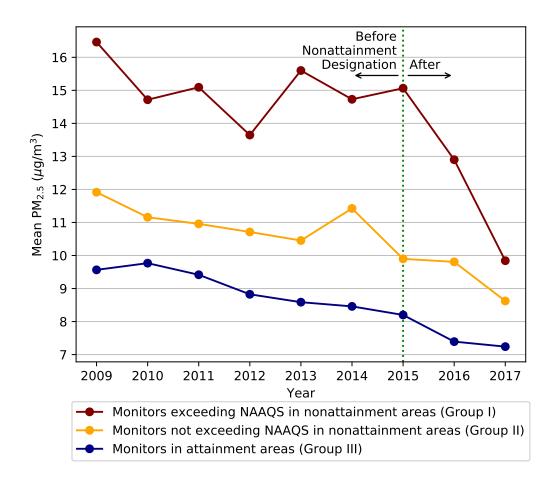


Figure 7: Average Monitor Readings by Attainment and NAAQS Status *Notes:* Monitor sample includes only monitors used for NAAQS compliance. Sample is restricted to monitors which began operation no later than 2013 and that operated at least through 2017. The red line shows the average of monitors that were in nonattainment areas and that were higher than the NAAQS in 2015 (Group I in the text). The yellow line shows the average of monitors that were in nonattainment areas and that were lower than the NAAQS in 2015 (Group II). The blue line shows the average of monitors in attainment areas (Group III).

Table 1: PM_{2.5} Monitor Counts by Frequency of Operation

	≤ 80 days	81–120 days	121–300 days	>300 days	Total
A. Monitors Operating in the Given Year					
2010	121	345	87	148	701
2011	95	341	55	177	668
2012	106	296	93	202	697
2013	83	313	117	229	742
2014	118	319	77	264	778
2015	112	381	59	290	842
2016	99	330	131	326	886
B. Monitors with 3 years of valid data for NAAQS assessment					
2013	72	274	73	157	576
2014	67	276	85	179	607
2015	90	290	62	215	657
2016	78	332	45	249	704
2017	77	299	96	274	746

Notes: Panel A reports the number of monitors designated as NAAQS primary compliance monitors which operated in the given year. Column 1 reports how many monitors operated no more than 80 days during that year, column 2 reports how many operated 81–120 days, and so on. Panel B reports how many monitors had sufficient data over the past three years that a design value could be calculated using that monitor. For example, a monitor that operated in 2016 but not 2015 would not be counted in 2016 while a monitor that operated in 2013–2015 would be counted in 2016.

Table 2: Misclassified Population by State

	Counties with no monitor	Counties with at least 1 monitor	Total
West Virginia	0	24,069	24,069
Tennessee	0	54,181	54,181
Arizona	0	195,751	195,751
Missouri	0	319,294	319,294
Kentucky	975,135	233,242	1,208,377
Pennsylvania	633,269	1,081,820	1,715,089
Ohio	945,497	1,240,213	2,185,710
Indiana	616,795	2,229,834	2,846,629
Texas	418,007	4,092,459	4,510,466
California	844,427	4,059,633	4,904,060
Illinois	6,437,475	0	6,437,475
Total	10,870,605	13,530,496	24,401,101

Notes: All misclassified counties in Illinois are counted as having no monitor because no monitor data were used in making attainment determinations in that state due to the monitors being deemed insufficiently accurate.

Table 3: Demographics by Attainment Status

	0 1	Misclassified	
Population (millions)	258.97	24.38	23.21
Share Urban	0.78	0.94	0.97
Race/Ethnicity			
White	0.67	0.54	0.39
Black	0.13	0.15	0.09
Hispanic	0.14	0.24	0.40
Asian	0.04	0.05	0.11
Other	0.09	0.13	0.23
Household Income			
<\$35,000	0.34	0.33	0.32
\$35,000–75,000	0.33	0.32	0.31
>\$75,000	0.33	0.34	0.37
Education			
No H.S. Diploma	0.14	0.16	0.21
H.S. Diploma	0.30	0.27	0.24
Some College	0.28	0.28	0.28
College Degree or More	0.28	0.29	0.27

Notes: Data for population, share urban, and race/ethnicity come from 2010 census block-level counts. Data for income and education come from 2005–2010 ACS block group–level estimates. Education sample is people age 25 and older. NAAQS limit is from the 2012 $PM_{2.5}$ rule and is $12\,\mu g/m^3$. Except for "fraction over \$75,000" and "fraction with some college," the difference in means between misclassified and nonattainment are all statistically significant with p-values less than 0.001.

Table 4: Effect of Nonattainment and NAAQS Status on Monitor Readings over Time

	(1)	(2)
Nonattainment×post×Over NAAQS		-2.3019***
		(0.5326)
Nonattainment×post	-1.1416***	-0.4729**
	(0.2435)	(0.2055)
2014	-0.1499***	-0.1508***
	(0.0432)	(0.0433)
2015	-0.4469***	-0.4478***
	(0.0472)	(0.0471)
2016	-1.1845***	-1.1851***
	(0.0509)	(0.0509)
2017	-1.4376***	-1.4382***
	(0.0599)	(0.0599)
\mathbb{R}^2	0.822	0.825
N	4712	4712

Notes: Outcome variable is annual average monitor reading of $\mu g/m^3$ PM $_{2.5}$. Regression sample includes all monitors used for NAAQS compliance, 2013–2017. The variable "post" is an indicator variable for years greater than 2015. "Nonattainment" is an indicator variable for monitors located in an area designated as nonattainment with the 2012 PM $_{2.5}$ standard. "Over NAAQS" is an indicator variable for monitors whose annual average in 2015 exceeded the NAAQS limit of 12 $\mu g/m^3$. Regressions also include monitor-level fixed effects. Standard errors clustered at the monitor level: ** p < .05, **** p < .01.

Supplementary Appendix

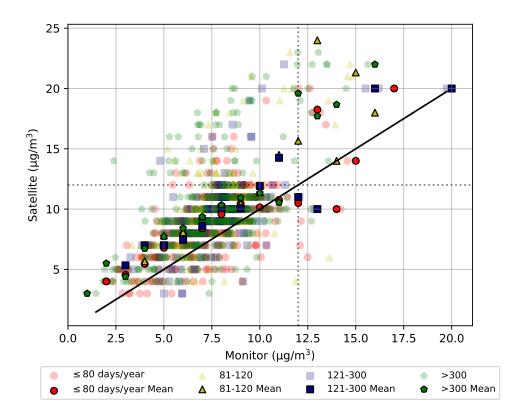


Figure A1: Globally Calibrated Satellite Readings versus Monitor Readings *Notes*: Horizontal axis is annual average monitor reading. Vertical axis is the satellite-derived reading for the 0.01-by-0.01-degree cell where the monitor is located. Red circles indicate monitors that operate no more than 80 days per year; yellow indicate triangles 81–120 days; blue squares indicate 121–300 days; and green pentagons indicate more than 300 days per year. Faint markers indicate individual marker–grid cell pairs; bold markers indicate the average for every bin centered at integers on the horizontal axis, i.e., satellite average for monitor readings of $1 \pm 0.5 \,\mu\text{g/m}^3$. Dashed gray lines show the $12 \,\mu\text{g/m}^3$ NAAQS threshold for nonattainment classification.

County	Core-based Statistical Area (Metro Area)
Yuma County	Yuma, AZ
Imperial County*	El Centro, CA
Kern County*	Bakersfield, CA
Los Angeles County*	Los Angeles-Long Beach-Anaheim, CA
Orange County*	Los Angeles-Long Beach-Anaheim, CA
Riverside County*	Riverside-San Bernardino-Ontario, CA
San Diego County	San Diego-Carlsbad, CA
Ventura County	Oxnard-Thousand Oaks-Ventura, CA
Cook County	Chicago-Naperville-Elgin, IL-IN-WI
Lake County	Chicago-Naperville-Elgin, IL-IN-WI
Madison County	St. Louis, MO-IL
St. Clair County	St. Louis, MO-IL
Bartholomew County	Columbus, IN
Cass County	Logansport, IN
Clark County	Louisville/Jefferson County, KY-IN
Floyd County	Louisville/Jefferson County, KY-IN
	Indianapolis-Carmel-Anderson, IN
	Seymour, IN
	Indianapolis-Carmel-Anderson, IN
	Chicago-Naperville-Elgin, IL-IN-WI
Marion County	Indianapolis-Carmel-Anderson, IN
Porter County	Chicago-Naperville-Elgin, IL-IN-WI
Shelby County	Indianapolis-Carmel-Anderson, IN
	Lafayette-West Lafayette, IN
	Evansville, IN-KY
	Terre Haute, IN
Bullitt County	Louisville/Jefferson County, KY-IN
Campbell County	Cincinnati, OH-KY-IN
Daviess County	Owensboro, KY
	Evansville, IN-KY
	Louisville/Jefferson County, KY-IN
Kenton County	Cincinnati, OH-KY-IN
St. Louis city	St. Louis, MO-IL
Butler County	Cincinnati, OH-KY-IN
Clermont County	Cincinnati, OH-KY-IN
Cuyahoga County*	Cleveland-Elyria, OH
Hamilton County	Cincinnati, OH-KY-IN
Jefferson County	Weirton-Steubenville, WV-OH
Montgomery County	Dayton, OH
Warren County	Cincinnati, OH-KY-IN
	Reading, PA
Berks County	Reading, 1 A
Berks County Cumberland County	Harrisburg-Carlisle, PA
	8
Cumberland County	Harrisburg-Carlisle, PA
Cumberland County Dauphin County	Harrisburg-Carlisle, PA Harrisburg-Carlisle, PA
Cumberland County Dauphin County Westmoreland County York County	Harrisburg-Carlisle, PA Harrisburg-Carlisle, PA Pittsburgh, PA
Cumberland County Dauphin County Westmoreland County	Harrisburg-Carlisle, PA Harrisburg-Carlisle, PA Pittsburgh, PA York-Hanover, PA Knoxville, TN
Cumberland County Dauphin County Westmoreland County York County Roane County Harris County	Harrisburg-Carlisle, PA Harrisburg-Carlisle, PA Pittsburgh, PA York-Hanover, PA Knoxville, TN
Cumberland County Dauphin County Westmoreland County York County Roane County Harris County Maverick County	Harrisburg-Carlisle, PA Harrisburg-Carlisle, PA Pittsburgh, PA York-Hanover, PA Knoxville, TN Houston-The Woodlands-Sugar Land, TX Eagle Pass, TX
Cumberland County Dauphin County Westmoreland County York County Roane County Harris County Maverick County Starr County	Harrisburg-Carlisle, PA Harrisburg-Carlisle, PA Pittsburgh, PA York-Hanover, PA Knoxville, TN Houston-The Woodlands-Sugar Land, TX Eagle Pass, TX Rio Grande City, TX
Cumberland County Dauphin County Westmoreland County York County Roane County Harris County Maverick County	Harrisburg-Carlisle, PA Harrisburg-Carlisle, PA Pittsburgh, PA York-Hanover, PA Knoxville, TN Houston-The Woodlands-Sugar Land, TX Eagle Pass, TX
	Yuma County Imperial County* Kern County* Los Angeles County* Orange County* Riverside County* San Diego County Ventura County Cook County Lake County Madison County St. Clair County Bartholomew County Clark County Hamilton County Johnson County Johnson County Johnson County Johnson County Warion County Shelby County Tippecanoe County Vanderburgh County Vigo County Bullitt County Campbell County Daviess County Henderson County Jefferson County St. Louis city Butler County Cuyahoga County Hamilton County Cuyahoga County Hamilton County Cuyahoga County Hamilton County Lefferson County Cuyahoga County Hamilton County Jefferson County Lefferson County Lefferson County Lefferson County Cuyahoga County Hamilton County Lefferson County

^{*} Partially misclassified county